

1 **Climate Change Stressors Destabilize the Microbiome of the Caribbean Barrel Sponge,**  
2 ***Xestospongia muta***

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16 **Running title**

17 Climate change effects on a sponge microbiome

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22 **Abstract**

23 The effect of climate change, both thermal stress and ocean acidification, on coral reefs is of  
24 increasing concern with the effects on calcification at the organismal level, and changes in the  
25 ratio of accretion to erosion on larger spatial scales of particular interest. But far fewer studies  
26 have been done on non-calcifying organisms, such as sponges, that have important ecological  
27 roles on coral reefs. Here we report the results of a combined thermal stress and ocean  
28 acidification experiment on the ecologically dominant barrel sponge, *Xestospongia muta*, found  
29 on coral reefs throughout the Caribbean basin. The results show that ocean acidification alone,  
30 as well as its interaction with elevated seawater temperature, has significant effects on the sponge  
31 microbiome. Specifically, the significant interactive effects of thermal stress and ocean  
32 acidification led to a decline in the productivity potential of the symbiotic cyanobacteria in these  
33 sponges with a subsequent impact on nutrient transfer, as carbohydrate, between symbiont and  
34 host. Additionally, while neither environmental stressor predictably changed sponge  
35 microbiome community composition, ocean acidification alone reduced the stability of sponge  
36 microbiomes and their predicted functions. Future changes in ocean acidification and thermal  
37 stress predicted by current climate models could negatively impact the microbiomes of coral reef  
38 organisms and therefore also affect their organismal performance and fitness in the future.

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40 **Keywords**

41 Sponges, microbiome, coral reefs, ocean acidification, climate change, Caribbean

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45 **1. Introduction**

46 The worldwide decline of coral reefs has been attributed to multiple environmental stressors but  
47 the consequences of climate change on coral reefs are widespread, long-term and potentially  
48 irreversible (Hoegh-Guldberg et al., 2007; Hough-Guldberg and Bruno, 2010). Both ocean  
49 acidification (OA) and elevated seawater temperatures (SWT) are known threats to coral reefs  
50 (Hoegh-Guldberg et al., 2007), with the interactive effects of increasing concern (Boyd et al.,  
51 2014; Pandolphi et al., 2011). Climate models estimate that 90% of the increased heat content of  
52 the atmosphere, as a result of the release of greenhouse gases (e.g., CO<sub>2</sub>), is sequestered in the  
53 oceans resulting in increasing seawater temperature (Donner et al., 2005) with multiple ecological  
54 effects including: species range extensions, increased incidence of opportunistic disease,  
55 enhanced storm effects, and changes in food-web structure (Doney et al., 2012). Simultaneously  
56 with increases in SWT, the increase in atmospheric CO<sub>2</sub> has caused a reduction of 0.1 pH units in  
57 the ocean, equivalent to a 30% decrease over pre-industrial era levels. Climate models predict  
58 additional declines of as much as 0.3-0.4 pH units by the year 2100 (Doney et al., 2012).  
59 However, these predictions are for the open ocean and there is now evidence that coastal marine  
60 environments, including coral reef ecosystems, currently exceed these predictions with significant  
61 temporal and spatial variability (Hofmann et al., 2011). A major ecological and economic  
62 concern is that lowering the pH and the aragonite saturation state [ $\Omega_{\text{arag}}$ ] will inhibit calcification  
63 in marine organisms (Ries et al., 2009).

64 Elevated SWT can cause coral bleaching, a stress response with significant impacts to coral  
65 reefs worldwide (Hoegh-Guldberg et al., 2007; Lesser, 2004). But fewer studies have considered  
66 the combined effects of elevated SWT and OA in combination (Anthony et al., 2011; Boyd et al.,  
67 2014), despite the fact that interacting stressors can be more damaging than the independent

68 effect of any single stressor (Byrne et al.,2013). Additionally, there have been far fewer studies  
69 of the effects of climate change stressors on non-calcifying taxa on coral reefs. For example,  
70 sponges are now recognized as a dominant taxon on many Caribbean coral reefs (Colvard et  
71 al.,2011), and it has been predicted that many coral reefs could become sponge dominated in the  
72 future (Bell et al.,2013). The ecological importance of sponges to coral reef communities is  
73 unequivocal (Bell et al.,2008)with sponges involved in a number of important roles on coral  
74 reefs including: benthic-pelagic coupling via filtration of large quantities of dissolved and  
75 particulate organic matter (de Goeij et al.,2013; Lesser and Slattery, 2013), nutrient cycling on  
76 coral reefs (Fiore et al., 2013 a, Southwell et al., 2008)and primary productivity and synthesis of  
77 secondary metabolites (Taylor et al.,2007). Many of the functional roles of sponges  
78 are dependent on diverse assemblages of symbiotic microorganisms(Erwin and Thacker, 2008;  
79 Fiore et al.,2010; Taylor et al.,2007).The barrel sponge, *Xestospongia muta*, is a long-lived and  
80 dominant component of the Caribbean benthic coral reef fauna (McMurray et al., 2010) whose  
81 multiple contributions to coral reef ecology and biogeochemistryhas been shown to be mediated  
82 by a diverse symbiotic microbiome (e.g., Fiore et al., 2013 a, b).Here, the results of an  
83 experiment examining the independent and interactive effects of SWT and OA on*X. muta*, and its  
84 symbiotic microbes, are presented.

85

## 86 **2. Materials and Methods**

### 87 *2.1. Study sites, experimental design and statistical analysis*

88 Barrel sponges(*Xestopongia muta*) were collected from South Perry Reef (17 m depth) on Lee  
89 Stocking Island in the Bahamas. All samples were collected whole by cutting the sponges from  
90 the substrate without compromising the spongocoel integrity and maintaining the pumping

91 activity of the sponges. Sponges ranged in size from 73-401 g wet weight. Sponges were  
92 randomly picked and acclimatized (N=5 for each treatment group) in ambient  $p\text{CO}_2$  and  
93 temperature reef seawater for 3 d within individual aquaria in a raceway system. Sponges were  
94 then acclimatized from the ambient conditions to experimental conditions over 2 d in a fully  
95 orthogonal matrix design of  $\text{CO}_2$  concentrations that reflected  $p\text{CO}_2$  (~390 ppm) conditions at  
96 the time of the experiments and predicted  $\text{CO}_2$  concentrations in the year 2100 under an A2  
97 climate model ( $p\text{CO}_2$  of ~800 ppm) scenario, as well as current and predicted summer SWT  
98 (IPCC, 2007). The seawater pH values for these conditions were held constant using the pH-stat  
99 approach with WTW 3310 pH meters (accuracy =  $\pm 0.005$  pH capability). The mean treatment  
100 temperatures were maintained using JBJ titanium heaters ( $\pm 0.5$  °C capability) and measured  
101 using HOBO Water Temperature Pro v2 data loggers (Onset Corp). Temperatures fluctuated on  
102 a diel basis in the flow through system as a result of daily heating and cooling, but still were  
103 within the 2-5.4°C predicted by the IPCC 2007 A2 model for the open ocean. The treatment  
104 groups were: Control (present day temperature and pH), Reduced pH (present day temperature  
105 and A2 pH), Elevated Temperature (A2 temperatures and present day pH) and Elevated  
106 Temperature and Reduced pH (A2 temperatures and pH). Sponges in all treatment groups were  
107 held in individual aquaria (8 L) with flowing seawater ( $1.0 \text{ L h}^{-1}$ ), and exposed to natural solar  
108 radiation under neutral density cloth for 12 d. The maximum irradiance of photosynthetically  
109 active radiation (PAR; 400-700 nm), measured with a LI-COR LI-192 cosine corrected  
110 underwater quantum sensor, was  $450\text{-}500 \mu\text{mol quanta m}^{-2} \text{ s}^{-1}$  and replicated the total irradiance  
111 of the site and depth of collection (e.g., Lesser and Gorbunov 2001). All treatment and time  
112 effects, and their interaction, were tested using ANOVA with *post hoc* multiple comparison  
113 testing (i.e., Tukey's HSD) as required.

114 2.2. Carbonate chemistry

115 Independent seawater samples (N=3) were collected from the experimental aquaria at the end of  
116 the experiment and analyzed for total alkalinity (TA;  $\mu\text{mol kg seawater}^{-1}$ ), total  $\text{CO}_2$  ( $\mu\text{mol kg}$   
117  $\text{seawater}^{-1}$ ) and the partial pressure of carbon dioxide ( $p\text{CO}_2$ ;  $\mu\text{atm}$ ) at the University of New  
118 Hampshire Ocean Process Analysis Laboratory (OPAL). TA was analyzed using an Apollo Sci-  
119 Tech AS-A2 automated analyzer, which employs the Gran titration procedure with a precision of  
120 0.1% and  $p\text{CO}_2$  ( $\pm 4 \mu\text{atm}$ ) using a LiCor 840A gas analyzer. The initial pH for the titration was  
121 measured on the sample using a Thermo Orion combination electrode (precision  $\pm 0.027$  pH  
122 units). Certified reference materials were used to ensure the precision of the measurements  
123 (Dickson et al., 2007).  $\text{pH}_T$  and  $\Omega \text{Ca}$  were then calculated with CO2calc software (Robbins et al.  
124 2010) using the temperature of each treatment, a salinity of 37.0 ppt, and the inorganic carbon  
125 dissociation constants from Mehrbach et al. (1973) as refitted by Dickson and Millero (2007).

126

127 2.3. Active fluorescence

128 The quantum yield of photosystem II (PSII) fluorescence was measured using a pulse amplitude  
129 modulated (PAM) diving fluorometer (Walz Inc.) at noon and midnight daily, equivalent to the  
130 maximum and minimum irradiances for each day of sampling, during the entire experiment. The  
131 *in vivo* fluorescence of chlorophyll varies between a minimum yield,  $F_o$ , and maximum yield,  $F_m$ .  
132 The difference between  $F_o$  and  $F_m$  fluorescent yields ( $F_m - F_o$ ) is the variable fluorescence,  $F_v$  and  
133 the ratio  $F_v/F_m$  or  $\Delta F/F_m'$  is the maximum (i.e., dark adapted) and steady state (i.e., in the light),  
134 respectively, quantum yield of PSII fluorescence (Warner et al., 2010). The PAM fluorometer  
135 utilized was a blue LED version (470 nm excitation, emission  $> 630$  nm) with the fiber-optic  
136 probe held 1 cm from the sponge in a perpendicular orientation. A saturation pulse of 0.8 s at

137 3000  $\mu\text{mol quanta m}^{-2} \text{ s}^{-1}$  was used to record maximal fluorescence ( $F_m$ ) at a gain setting of 6.  
138 While studies on cyanobacteria are generally conducted using red excitation, the cyanobacteria,  
139 *Synechococcus* sp., in *Xestospongia muta* contain phycoerythrin and as a result they absorb more  
140 efficiently in the blue portion of the spectrum. Additionally, the blue LED detects a wider  
141 emission signal ( $>630 \text{ nm}$  versus  $>700 \text{ nm}$  for the red LED PAM) making it more sensitive than  
142 the red LED version of the PAM. Lastly, as irradiance and pigment content has the greatest  
143 influence on cyanobacterial quantum yields (Campbell et al., 1998), and all sponge treatment  
144 irradiances and treatment pigment concentrations were not significantly different from each other  
145 (Lesser, unpublished data), these quantum yield measurements are a very good approximation of  
146 treatment effects on the quantum yield of cyanobacterial PSII fluorescence. All sponge  
147 measurements ( $N=3$ ) were taken at the same instrument settings. In cyanobacteria the quantum  
148 yields of PSII fluorescence, under constant pigment concentration and irradiances, also correlate  
149 very well with rates of photosynthesis (i.e., oxygen evolution) especially when the same sample  
150 is measured repeatedly as was done here (Campbell et al., 1998). All quantum yield  
151 measurements are ratios and therefore *a priori* not normally distributed. These measurements  
152 were transformed ( $\text{Log} +1$ ) prior to analysis and back transformed for presentation. Only the  
153 measurements at the end of the experiment were analyzed and reported herewith ocean  
154 acidification, SWT and time of day (i.e., midday or midnight; for each sampling time:  $N=20$   
155 from four treatment groups and five sponges) each treated as an independent factor using a three-  
156 factor ANOVA with interaction for the analysis of the data.

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159

160 2.4. Metagenetic analysis of 16S rRNA genes

161 For the metagenetic analysis of 16S rRNA genes, DNA extractions were performed on the  
162 experimental sponge samples (N=3 for each treatment) at the end of the experiment as described  
163 by Fiore et al. (2013 a). A section of each sponge, including both the pinacoderm and outer  
164 mesohyl of the sponge, was cut into smaller pieces for processing. The 16S rRNA genes of each  
165 sample was amplified and barcoded for multiplexed pyrosequencing using Titanium adapter  
166 sequences A (forward primer) and B (reverse primer), and a 10 bp barcode sequence added to the  
167 PCR primers. Primers designed to amplify both Bacteria and Archaea (hypervariable V6 region)  
168 were used, consisting of the forward primer U789F (5'-TAGATACCCSSGTAGTCC-3') and the  
169 reverse primer U1068R ('-CTGACGRRCGCCATGC-3'). Samples were pyrosequenced on a  
170 ROCHE/454 GS FLX+ platform (Roche, Branford, CT, USA) at the University of Illinois W.M.  
171 Keck Center for Comparative and Functional Genomics (Urbana-Champaign, IL, USA). The  
172 sequence analysis is similar to that described in Fiore *et al.* (2013) and uses the Quantitative  
173 Insights Into Microbial Ecology (QIIME) pipeline (Caporaso et al., 2010) on the Amazon Elastic  
174 Compute Cloud (EC2), except where noted. Raw sequence reads were filtered for quality by  
175 discarding short reads (<200 bp), or reads with more than two mismatches with the primer  
176 sequence, or with ambiguous nucleotides, or with an average quality score less than 25. A  
177 custom Perl script based on the QIIME script "split\_libraries.py", was used to trim primers from  
178 the sequences, assign reads to their sample of origin (based on MID tags), and reverse  
179 complement the reads originating from the B adapter (reverse reads). Trie clustering (QIIME  
180 team, unpublished, <http://qiime.org>) was used to collapse reads that are prefixes of each other  
181 into clusters and discard singleton reads as described in Fiore et al. (2013). Chimeric sequences  
182 were identified and removed using USEARCH 6.1 (Edgar 2010) in QIIME. Reads were then



183 clustered into OTUs using UCLUST de novo clustering (97% similarity) (Edgar 2010).  
184 Taxonomy was assigned to representative sequences for each OTU using the Ribosomal  
185 Database Project (RDP) classifier with a minimum cutoff of 0.8 (Wang et al.,2007) in QIIME.  
186 The OTU table generated in QIIME was rarefied (min=4331), and any OTUs with less than  
187 4,331 sequences were removed prior to downstream analysis. ANOSIM was utilized to examine  
188 treatment effects on the community composition of experimental sponges. Analysis of microbial  
189 or functional  $\beta$ -diversity was quantified using weighted UniFrac distances for microbial  
190 communities (a quantitative, phylogenetic  $\beta$ -diversity metric) or Bray-Curtis divergences for  
191 functional profiles (a quantitative, non-phylogenetic measure) in each individual sponge. These  
192 metrics are then used to assess the stability of the individual sponge samples as it relates to any  
193 treatment effects. The significance of differences between all pH treatments, all A2 treatments,  
194 or all treatment categories was assessed using permutational *t*-tests of distances, with a  
195 Bonferroni correction applied to comparisons between treatments to account for multiple  
196 comparisons. Sequences resulting from pyrosequencing were deposited in the iMicrobe data  
197 repository under project accession number CAM\_P\_0000957  
198 (<ftp://ftp.imicrobe.us/projects/103/samples/>).

199

## 200 *2.5. Functional analysis of sponge microbiomes using PICRUSt*

201 Functional profiles for 16S rRNA gene sequence data were predicted using the program  
202 Phylogenetic Investigation of Communities by Reconstruction of Unobserved States (PICRUSt)  
203 version 1.0 (Langille et al.,2013). PICRUSt uses an algorithm that estimates the functional gene  
204 content of Bacteria or Archaea for which no genome is available. PICRUSt does this through  
205 evolutionary modeling of the copy number of each gene family, based on the strain's

206 phylogenetic relationship with all Bacteria and Archaea for which sequenced genomes are  
207 available. The phylogenetic predictability of sponge samples was calculated using Nearest  
208 Sequenced Taxon Index (NSTI) scores, which measure the average branch length on the  
209 reference phylogeny (Greengenes version 13.8) between the representative sequence for each  
210 OTU in a sample compared against a sequenced genome, weighted by the abundance of that  
211 OTU.

212

### 213 *2.6. Proximate biochemical composition*

214 The proximate biochemical composition of *Xestospongia muta* was determined from lyophilized  
215 tissue (including host and microbial symbionts) samples of replicate sponges (N=5) from each  
216 treatment. For each sponge, freeze-dried “tissue” was ground into a fine powder using a Wiley  
217 Mill. Levels of soluble protein and soluble carbohydrate were determined colorimetrically using  
218 the Bradford(1976) and Dubois et al. (1953) techniques, respectively as described by Slattery and  
219 McClintock (1995). Lipid was measured gravimetrically using the technique of Freeman et  
220 al.(1957), and ash was determined by placing tissues in a muffle furnace for 4 h at 500°C(Paine,  
221 1971). The remaining refractory material was calculated by subtraction and considered to be  
222 insoluble protein(Lawrence, 1973).

223

## 224 **3. Results**

### 225 *3.1. Carbonate chemistry*

226 Sponges were acclimatized to either current levels of OA ( $p\text{CO}_2 \sim 390\text{ppm}$ ; resulting pH:  $8.07 \pm$   
227  $0.05$  [SD]) or predicted pH levels for the year 2100 ( $p\text{CO}_2 = 800\text{ppm}$ ; resulting pH:  $7.81 \pm 0.12$   
228 [SD]) under an A2 climate model(IPCC 2007). Within each  $p\text{CO}_2$  treatment, sponges were also

229 acclimatized to current temperatures (mean  $29.2 \pm 0.9$  [SD] $^{\circ}\text{C}$ ; Range 27.3-30.1 $^{\circ}\text{C}$ ), or elevated  
230 temperatures (mean  $31.4 \pm 1.07$  [SD] $^{\circ}\text{C}$ ; Range 29.3-33.4 $^{\circ}\text{C}$ ) representative of predicted  
231 summertime means for the year 2100(IPCC 2007) along with the daily variability that  
232 characterizes coastal environments. The mean differences in treatment conditions were significant  
233 for both OA (*t*-test; two-tailed,  $P < 0.0001$ ) and temperature (*t*-test; two-tailed,  $P < 0.0001$ ). All  
234 measured and predicted carbonate chemistry values for each treatment group are presented in  
235 Table 1. These predictions were based on the IPCC 2007 Coupled Model Intercomparison  
236 Project that have now been replaced with new climate model scenarios, the Representative  
237 Concentration Pathways (RCP) from the IPCC 2014 (Rogelj et al., 2012) report. The A2 climate  
238 model conditions used in these experiments were closest to the RCP6.0 climate model when you  
239 include the  $\text{CO}_2$  equivalents for  $\text{CH}_4$  and  $\text{N}_2\text{O}$  (Rogelj et al., 2012).

240

### 241 *3.2. Active fluorescence*

242 The effects of SWT and OA on the photosynthetic symbionts were assessed using both the  
243 maximum ( $F_v/F_m$ ) and steady state ( $\Delta F/F_m'$ ) quantum yields of photosystem II (PSII)  
244 fluorescence measured at noon and midnight with a pulse amplitude modulated (PAM)  
245 fluorometry (Fig. 1). No visible signs of “bleaching” were observed in experimental sponges.  
246 Because the interaction of temperature, pH and time on the quantum yields of PSII fluorescence  
247 was significant (ANOVA:  $F=9.6$ ,  $P=0.004$ ), no further analysis of any independent effects was  
248 considered. A Tukey’s HSD multiple comparison test on the complete matrix of treatment  
249 groups reveals that significant decreases in the quantum yields of PSII fluorescence were  
250 observed (Fig. 1) in the reduced pH and elevated temperature treatments, with no differences  
251 between noon and midnight measurements, when compared to control values. However,

252 the combined elevated temperature and reduced pH treatments ( $0.426 \pm 0.014$  [SE]) had the  
253 lowest quantum yields observed compared to the control treatment group ( $0.621 \pm 0.018$  [SE]).

254

### 255 3.3. Metagenetic analysis of 16S rRNA

256 Pyrosequencing of the 16S rRNA gene for *Xestospongia muta* symbionts yielded 282,298 reads  
257 (average read length 299 nucleotides). Following quality filter steps and removal of singleton  
258 reads there were 206,995 reads that were then clustered into OTUs at 97% similarity. A total of  
259 967 OTUs remained following removal of chimeric sequences, singletons and contaminants from  
260 the sponge samples, and these included 19 known bacterial and two archaeal phyla (Fig. 2). Two  
261 groups of unclassified prokaryotes were clustered as “Unclassified Bacteria” and “Unclassified  
262 Archaea” (Fig. 2). These results are similar to those for *X. muta* from other locations around Lee  
263 Stocking Island, as well as from populations in the Cayman Islands and Florida Keys (Fiore et  
264 al., 2013 a). An analysis of similarity (ANOSIM) revealed no significant differences between  
265 treatments for the sponge microbial communities (Global  $R=0.005$ ,  $P=0.31$ ).

266

### 267 3.4. Functional analysis of sponge microbiomes using PICRUSt

268 To assess whether functional changes in the symbiotic communities occurred in *Xestospongia*  
269 *muta* from each treatment group, functional profiles were calculated using the PICRUSt software  
270 package. This approach predicts the gene families present in microbial metagenomes by  
271 modeling the gene families present in each OTU using available fully sequenced genomes as a  
272 reference (Langille et al., 2013). No significant differences were found between predicted  
273 functional categories by PICRUSt. Testing of whether any functional categories were enriched  
274 was done by summarizing the PICRUSt-predicted gene family counts (i.e., KEGG orthology or

275 KO counts), and no significant differences were found by Kruskal-Wallis tests of KEGG  
276 Pathway abundance, summarized at the second or third levels of the KEGG functional hierarchy  
277 ( $p > 0.05$ , FDR  $q > 0.05$ ). Sponge samples were also tested for differences in the abundance of  
278 specific functional categories (N=264) across all treatments by g-test of predicted results, rarified  
279 to 20,000 counts per sample. The top-three categories were all genes involved in photosynthesis:  
280 'photosynthesis proteins', 'photosynthesis', and 'photosynthesis – antennae proteins'. Using a  
281 false discovery rate control for multiple comparisons, only the first of these categories  
282 was significant (FDR  $q = 0.14$ ,  $p = 0.0005$ ). The mean relative abundance for photosynthesis  
283 proteins, after rarefaction, was lowest in the elevated temperature and reduced pH treatment  
284 group.

285 In order to test the potential accuracy of phylogenetic prediction of functional profiles in  
286 these samples, the average phylogenetic distance between microbial OTUs in sampled sponges  
287 versus the nearest sequenced microbial genome was calculated. These distances were weighted  
288 by the abundance of each OTU in the sample using the Nearest Sequenced Taxon Index, or NSTI  
289 (Langille et al., 2013). The mean NSTI score was 0.26, with values ranging from 0.11 to 0.31  
290 (Fig. 3), and were lowest when reference genome coverage was best represented in the  
291 characterized microbiome (Fig. 2). A two-factor ANOVA of the independent and interactive  
292 treatment effects on the Log +1 transformed NSTI values shows no treatment effects were  
293 detected (ANOVA:  $P > 0.05$ ).

294 Finally, a test of whether OA or elevated SWT altered the extent of inter-individual  
295 variation in the sponge microbiome or its predicted function was conducted. Within each  
296 treatment, changes in sample-to-sample turnover in microbial community composition, as a  
297 measure of community stability (i.e.,  $\beta$ -diversity), were assessed using Weighted UniFrac

298 distances, and changes in predicted function were assessed using Bray-Curtis divergences. The  
299 resulting  $\beta$ -diversity values were compared for either all reduced pH treatments (Fig.4a,b), all  
300 elevated temperature treatments (Fig.4c,d), or each treatment individually (Fig.4e,f). Despite the  
301 absence of significant changes in community composition (see Fig. 1), reduced pH alone  
302 measurably increased the inter-individual variation in sponge microbiomes, while the elevated  
303 temperature alone treatment reduced this variation (Fig. 4).

304

### 305 *3.5. Changes in the proximate biochemical composition of sponges*

306 Quantifying the proximate biochemical composition (i.e., protein, lipid and carbohydrate)of the  
307 holobiont at the end of the experiment revealed no significant effects of OA, SWT or their  
308 interaction on the concentration of protein or lipids (Fig. 5). The concentration of carbohydrates,  
309 however, showed a significant interactive effect of reduced pH and elevated  
310 temperatures(ANOVA:  $F=18.8$ ,  $P=0.0005$ ). Post-hoc multiple comparison tests (i.e., Tukey's  
311 HSD) shows that the A2predicted temperature and pH treatment group (i.e., Elevated  
312 temperature and reduced pH) had significantly ( $P<0.05$ ) lower concentrations of carbohydrate  
313 when compared to all other treatment groups (Fig. 5).

314

## 315 **4. Discussion**

316 The interactive effects of elevated SWT and OA are of increasing concern in marine ecosystems  
317 generally (Boyd et al.,2014), and on coral reefs specifically (Hoegh-Guldberg and Bruno, 2010).  
318 In addition to scleractinian corals, spongescan also be negatively affected by the impacts of  
319 environmental stressors, including elevated SWT (Fan et al.,2013; Webster et al.,2013)  
320 orOA(Goodwin et al.,2013). Alternatively, sponges can also benefit from the effects of

321 increasing CO<sub>2</sub>; studies done at natural CO<sub>2</sub>seeps that simulate future OA conditions,in the  
322 absence of elevated SWT, show that sponges with photoautotrophic symbionts (e.g.,  
323 *Synechococcus* sp.) increase in abundance compared to adjacent sites with contemporary  
324 carbonate chemistry (Morrow et al.,2014).This is consistent with studies on free-living  
325 *Synechococcus* that showed an increase in their rates of photosynthesis and cell division (Fu et  
326 al.,2007)when exposed to similar conditions to those described here for the A2 predicted  
327 temperature and pH treatment group.While the symbiotic cyanobacteria of sponges can supply a  
328 significant amount of carbon to their host via autotrophy (Wilkinson, 1983), cyanobacteria can  
329 also be negatively affected by OA conditions under nutrient limitation (e.g., Shi et al.,2012), a  
330 situation unlikely to occur *in hospite*for sponges (Fiore et al., 2010).

331 Cyanobacteria in the genus *Synechococcus* are well-described symbionts of *Xestospongia*  
332 *muta*(Fiore et al., 2013 a) and a previous study has shown that *X. muta* harbors different  
333 symbiotic phylotypes of *Synechococcus spongiarum* (Erwin and Thacker, 2008 a).  
334 Observations on other sponge species(Erwin and Thacker, 2008 b) suggest that different  
335 phylotypes of cyanobacteria in *X. muta*may confer varying abilities to translocate  
336 photoautotrophically derived organic products to the host. One study on *X. muta*, using isotopic  
337 tracers (e.g., NaH<sup>13</sup>CO<sub>3</sub>), shows that the bacterial community does readily fix carbon and  
338 translocate labeled products to the host (Fiore et al., 2013 b).

339 Whenthe symbiotic cyanobacteria of sponges are exposed to elevated *p*CO<sub>2</sub>alone their  
340 abundance increasessignificantly with the most important predicted functional role of these  
341 symbionts being photosynthesis (Morrow et al., 2014), while elevated SWT alone has been  
342 shown to significantly decrease steady state quantum yields of PSII fluorescence in sponges with  
343 cyanobacteria (Cebrian et al., 2011). The results presented here show that the interactive effects

344 of OA and elevated SWT affect the photosynthetic apparatus of the symbiotic cyanobacteria in  
345 *Xestospongia muta* by decreasing the number of functional PSII units. Additionally, the sponges  
346 in this experiment not only experienced the mean predicted future values for the A2 climate  
347 model, but also the diel variability naturally experienced on shallow coral reefs (e.g., Hoffmann  
348 et al., 2011) where the extreme values can actually exceed the predicted A2 predicted  
349 climatology. Under these experimental conditions, *X. muta* showed a decrease in both the steady  
350 state and maximum quantum yields of PSII fluorescence. This is the result of either non-  
351 photochemical quenching with state transition changes as the underlying mechanism (Campbell  
352 et al., 1998), or chronic photoinhibition with damage to PSII that would result in a decrease in  
353 productivity (Gorbunov et al., 2001). While state transitions could potentially be affecting these  
354 measurements, both dark-adapted and effective quantum yields of PSII fluorescence show the  
355 same patterns. Since state transitions largely occur during darkacclimation (Campbell et al.,  
356 1998), if one compares maximum and effective yields state transitions appear to have had very  
357 little effect on the treatment specific patterns which are the same for both fluorescent  
358 measurements. These decreases in PSII quantum yields would then be mechanistically linked to  
359 lower rates of photosynthesis in cyanobacteria (Campbell et al., 1998), and the observed decrease  
360 in carbohydrate concentration in the holobiont.

361 Other studies have shown the effects of both elevated SWT and OA on sponges  
362 including decreases in the biosynthesis of secondary metabolites known to be involved in  
363 chemical defense (Duckworth et al., 2012). Additionally, studies on the excavating sponge  
364 *Cliona orientalis* have shown that the interactive effects of OA and elevated SWT can negatively  
365 affect energetic budgets, but sponge growth and reef bioerosion would still increase (Fang et  
366 al., 2013, 2014). Experiments using a multiplexed qPCR approach with sponge explants exposed



367 to elevated SWT have shown significant effects on host function as sponges became necrotic  
368 during the experiment (Fanet al., 2013). In the same experiment the microbiome community of  
369 necrotic sponge explants was significantly different from other treatment groups as were the  
370 metagenomic and metaproteomic profiles (Fanet al., 2013).

371 The experiment presented here, using intact whole sponges, provides a more ecologically  
372 realistic scenario on the effects of multiple stressors occurring simultaneously on this  
373 sponge. While these physiological changes were not accompanied by significant community  
374 changes in the microbiome of *Xestospongia muta*, predicted OA, but not elevated SWT,  
375 significantly increased the inter-individual variation in the composition (Fig. 4a,c,e) and predicted  
376 function (Fig. 4b,d,f) of the *X. muta* microbiome. This may reflect a decreased ability of the host  
377 to regulate its microbiome under stressful conditions. Along with this destabilization of sponge  
378 microbiomes, significant changes in the predicted abundance of microbial 'photosynthesis  
379 proteins' were observed using PICRUSt. This decrease in photosynthesis proteins is consistent  
380 with the significant decrease in the quantum yields of PSII fluorescence for the cyanobacterial  
381 symbionts of *X. muta* in the A2 predicted temperature and pH treatment group. PICRUSt is  
382 dependent on the availability of reference genomes to translate microbial taxonomy into  
383 predicted function, and the NSTI values observed here are 2-3 times greater (~0.24 to 0.29 across  
384 treatments, Fig. 3) than the gut microbiome of diverse terrestrial mammals but comparable with  
385 other underexplored environments such as the Guerrero Negro microbial mats (NSTI = 0.23) for  
386 which insufficient genomic resources are available (Langille et al., 2013). Thus current function  
387 predictions in sponges should serve to highlight the need for targeted cultivation and sequencing  
388 of underexplored lineages in the sponge microbiome in order to improve the accuracy of future  
389 predictions.

390 While it is true that genomic information for the symbionts of sponges and many other  
391 marine invertebrates is generally lacking, genome databases are populated with much genomic  
392 data for two key photosymbionts detected in these samples: the genus *Synechococcus* sp.,  
393 including *Candidatus Synechococcus spongiarum*, a well described cyanobacterial symbiont of  
394 sponges (Burgsdorf et al., 2015) and Chloroflexi, a group of photoheterotrophic green non-sulfur  
395 bacteria that is tolerant of both elevated temperatures and lower pH. The latter increased by ~20%  
396 in both treatments exposed to elevated temperatures (Fig. 2). Chloroflexi are often found in  
397 syntrophic relationships with cyanobacteria, and the extensive literature on their roles in  
398 microbial mat physiology should inform future studies on their functional roles in sponges (Klatt  
399 et al., 2013).

400

## 401 **5. Conclusions**

402 Experiments simulating coastal environments, including coral reefs, are important  
403 since increasing SWT and OA predicted by climate models for open ocean ecosystems in the year  
404 2100 already impacts these nearshore ecosystems. The data presented here highlight how much  
405 remains to be discovered about the physiology of sponges and their microbiomes under present or  
406 predicted climate change conditions. The microbiome of *Xestospongia muta* is clearly less  
407 represented in the available genomes than many other organisms with microbial symbionts or  
408 unexplored environments (Langille et al., 2013) which currently limits the predictive power to  
409 quantify their function, and changes in function, under predicted climate change scenarios using  
410 programs like PICRUSt. PICRUSt, however, has been used successfully here to show our best  
411 current estimate of the functional consequences of changes in sponge microbiomes based on  
412 available genomic data. A similar approach produced ecologically plausible predictions for a

413 con-specific sponge, *Xestospongia testudinaria*, with similar NSTI scores (de Voogd et al.,  
414 2015). As discussed in previous works (de Voogd et al., 2015) comparisons of PICRUSt  
415 accuracy using paired sponge samples deeply sequenced using both 16S rRNA amplicons and  
416 shotgun metagenomics is needed in order to quantify whether PICRUSt's prediction accuracy in  
417 sponges is more like soils (which achieved high accuracy despite NSTIs ~0.17), or more like the  
418 Guerrero Negro hypersaline microbial mats.

419       Regardless, the results presented here indicate that *X. muta* and its microbiome  
420 responds negatively to the independent effects of OA, and the interactive effects of elevated SWT  
421 and OA at the physiological and organismal levels. The ecological consequences (e.g., decrease  
422 in fitness), if any, for the effects of climate change related stressors are not currently known for  
423 this ecologically important member of coral reef communities throughout the Caribbean basin  
424 (McMurray et al., 2010). Recent studies have clearly shown the importance of metabolic  
425 interchange between the host and symbionts of *X. muta* (Fiore et al., 2015) such that  
426 understanding the effects of predicted climate change on the physiology of *X. muta* and their  
427 symbionts is important, and how those effects translate to the level of reef communities is  
428 essential.

429

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597  
598

599 **Figure Legends**

600 Figure 1. Treatment effects on the maximum ( $F_v/F_m$ , grey bars [midnight]) and steady state  
601 ( $\Delta F/F_m'$ , white bars [midday]) quantum yields of photosystem II (PSII) fluorescence (mean  $\pm$   
602 SE) for *Xestospongia muta*. Superscripts denote groups not statistically different from one  
603 another using multiple comparison testing (Tukey's HSD).

604

605 Figure 2. Average relative abundance of OTUs (97% similarity) at the class level for each  
606 treatment group for *Xestospongia muta* included unclassified Bacteria, unclassified Archaea,  
607 Crenarchaeota, Euryarchaeota, Acidobacteria, Actinobacteria, AncK6, Bacteroidetes,  
608 Chloroflexi, Cyanobacteria, Firmicutes, Gemmatimonadetes, H178, Nitrospirae, OD1,  
609 PAUC34f, Poribacteria, Proteobacteria, SRB1093, Spirochaetes, TM7, Thermi and  
610 Verrucomicrobia.

611

612 Figure 3. The Nearest Sequenced Taxon Index (NSTI, mean  $\pm$  SD) for all treatment groups.  
613 Samples with more organisms closely related to those with sequenced genomes have lower NSTI  
614 scores.

615

616 Figure 4. Effects of pH and thermal stress on the structure and function of *Xestospongia muta*  
617 microbiomes. Box plots show the effects of predicted (IPCC 2007 A2 scenario) pH and  
618 temperatures on sponge microbiomes. Phylogenetic  $\beta$ -diversity was measured using the  
619 weighted UniFrac metric, while predicted functional  $\beta$ -diversity was assessed using the Bray-  
620 Curtis distance between PICRUSt predicted functional profiles for each sample. Significance  
621 between conditions was assessed by Bonferroni-corrected permutational *t*-tests. All pairwise

622 differences were tested, only those with Bonferroni-corrected  $p < 0.10$  are shown. (A, B) Effects  
623 of pH across all temperature regimes (i.e. Control and Elevated Temperature treatments vs.  
624 Reduced pH and Reduced pH and Elevated Temperature treatments). (C, D) Effects of  
625 temperature across all pH regimes. (E, F) separate and combined effects of pH and temperature.  
626

627 Figure. 5. Proximate biochemical composition (mean  $\pm$  SE) of *Xestospongia muta* ( $\mu\text{g mg}^{-1}$   
628 tissue) for each treatment group. Superscripts denote groups not statistically different ( $P > 0.05$ )  
629 from one another using multiple comparison testing (Tukey's HSD).  
630

Table 1. Values for directly measured and calculated carbonate chemistry parameters for the experimental treatments. Parameters of carbonate seawater chemistry were calculated from TA, TCO<sub>2</sub>, pCO<sub>2</sub>, temperature and salinity using the free-access CO2Calc package.

<b>Treatment</b>	<b>TA (<math>\pm</math> SD)*</b> ( $\mu\text{mol kg}^{-1}$ SW)	<b>TCO<sub>2</sub> (<math>\pm</math> SD)*</b> ( $\mu\text{mol kg}^{-1}$ SW)	<b>pCO<sub>2</sub> (<math>\pm</math> SD)*</b> ( $\mu\text{atm}$ )	<b>pH<sub>T</sub>#</b>	<b><math>\Omega_{\text{arg}}</math>#</b>
Control	2376 $\pm$ 26	2019 $\pm$ 32	408	8.035	4.02
Reduced pH	2403 $\pm$ 48	2162 $\pm$ 44	717	7.841	2.86
Elevated Temperature	2376 $\pm$ 19	2019 $\pm$ 27	444	8.005	4.15
Elevated temperature and Reduced pH	2403 $\pm$ 52	2162 $\pm$ 53	861	7.772	2.97

\* -directly measured

# -calculated

SW -seawater

pH<sub>T</sub> -total scale

FIGURE 1

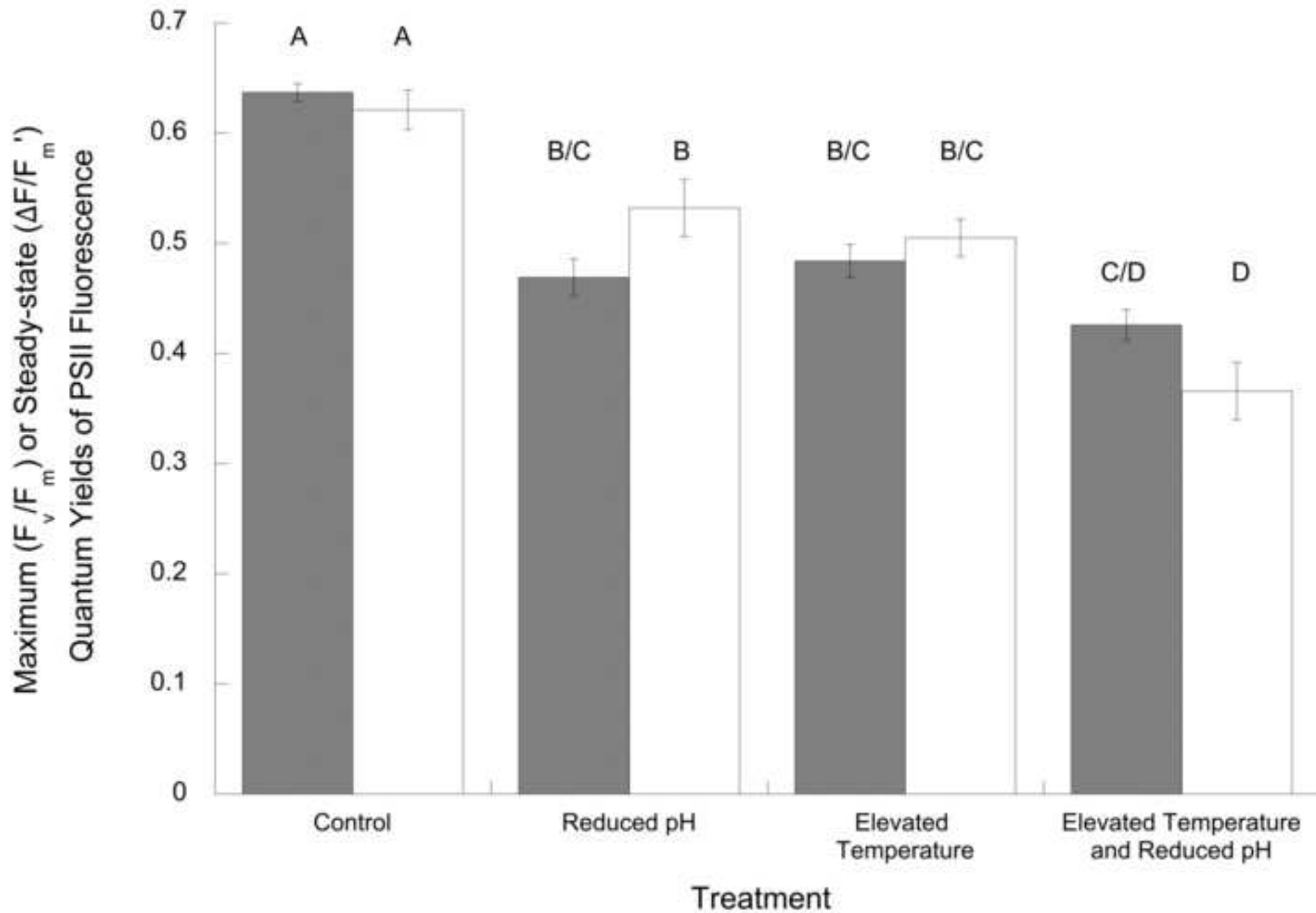


FIGURE 2

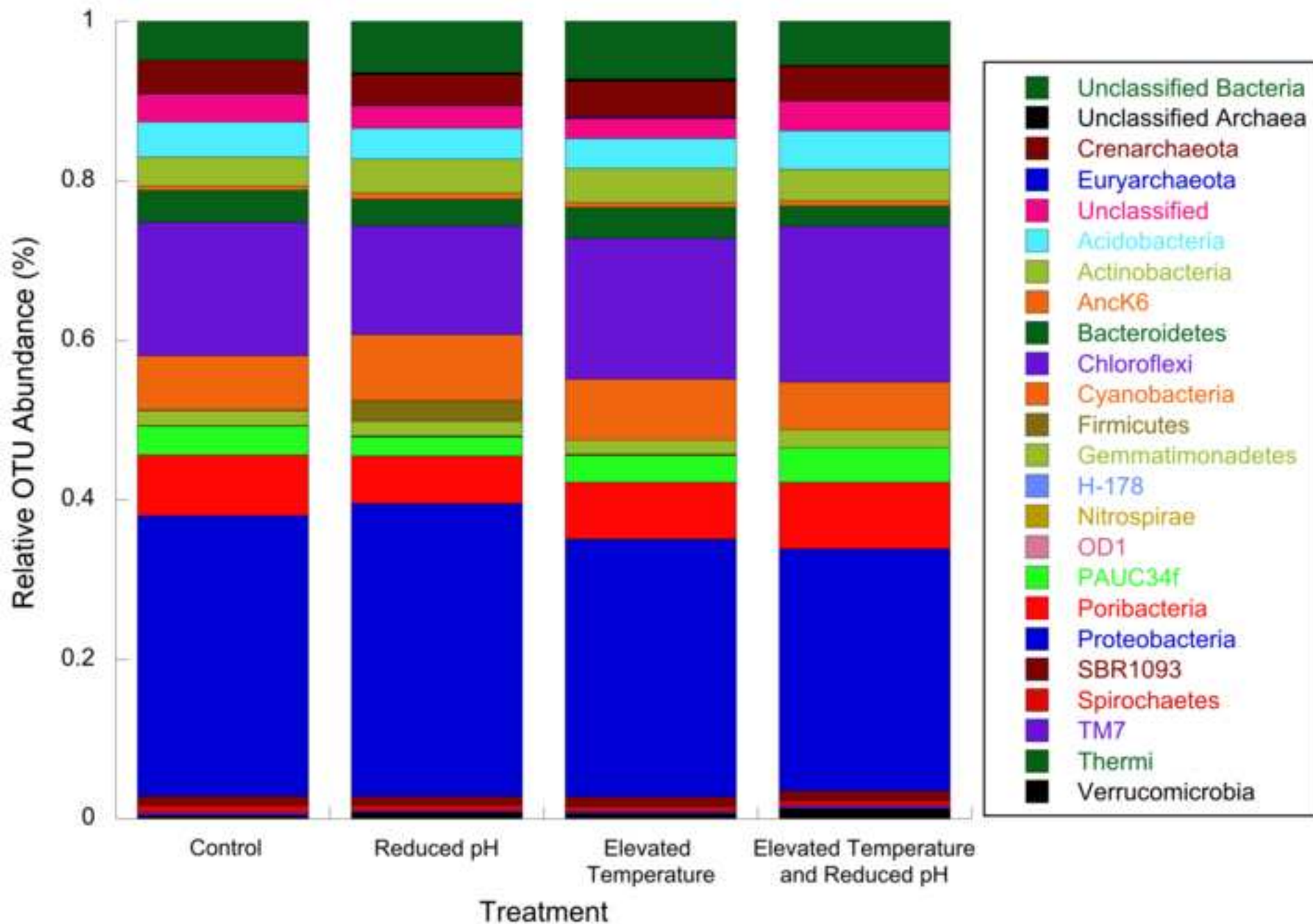




FIGURE 3

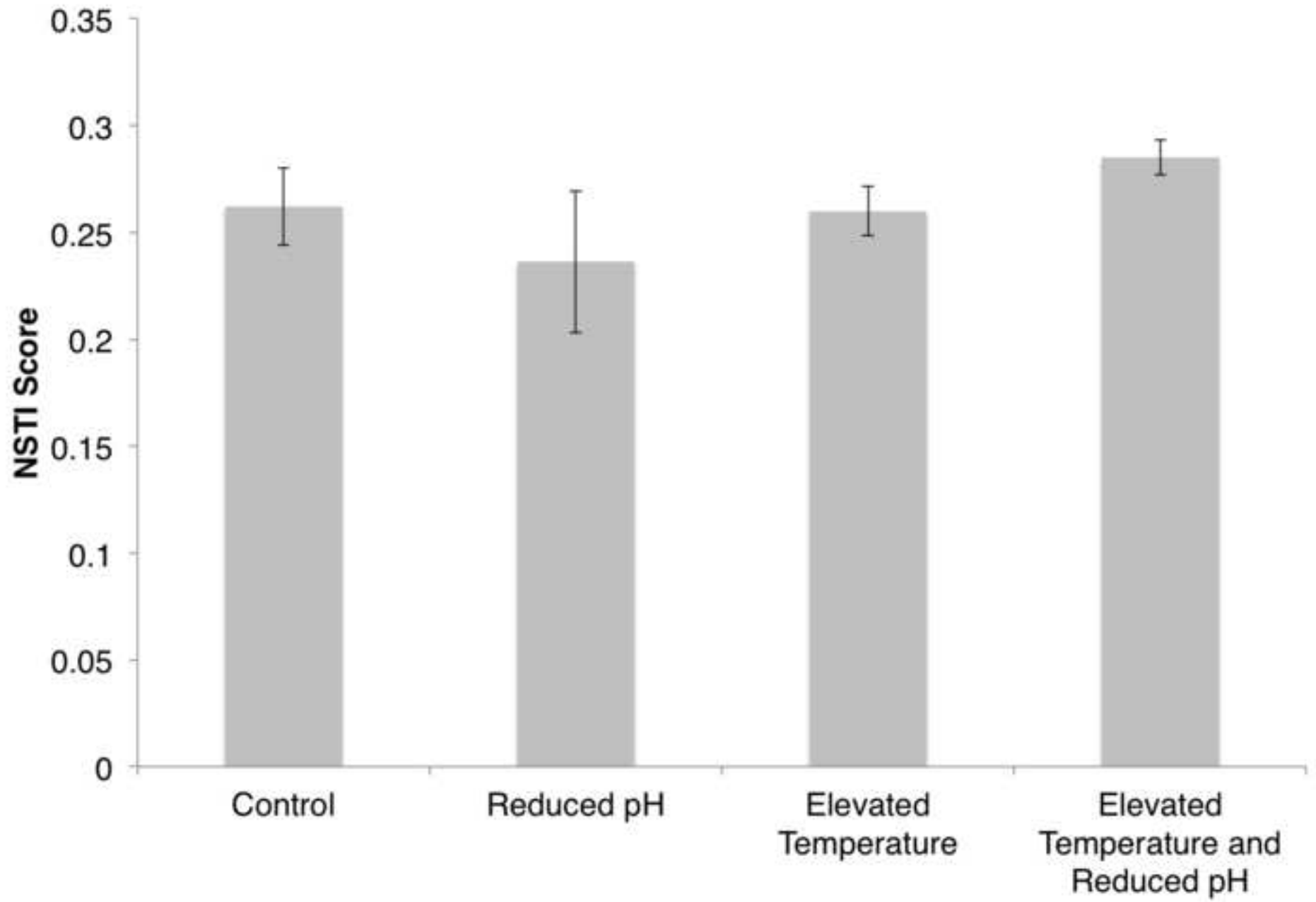


FIGURE 4

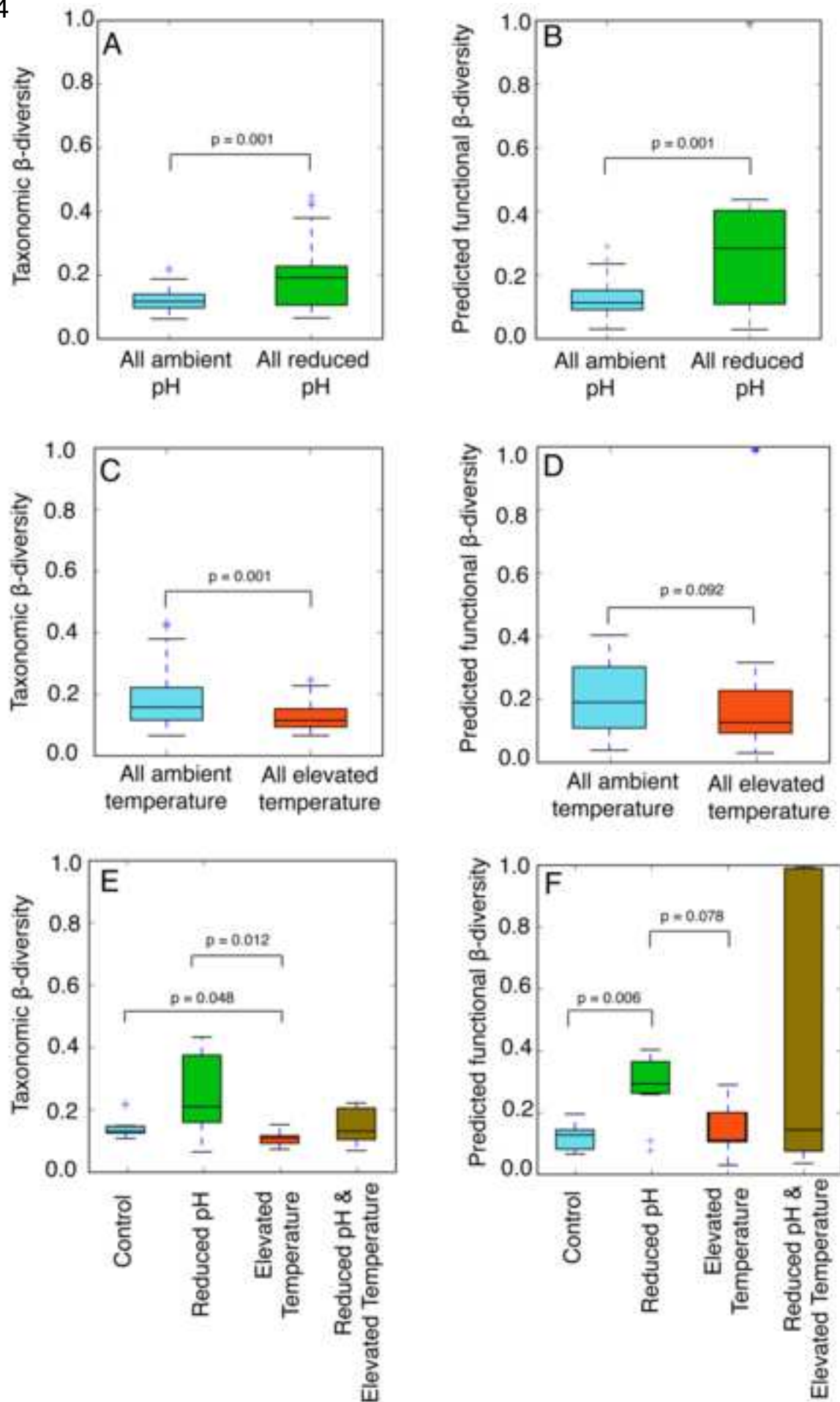


FIGURE 5

